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[Long-Term Fertiliser Regimes have both Direct and Indirect Effects on Arthropod Community Composition and Feeding Guilds.](#)

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1 **Long-Term Fertiliser Regimes have both Direct and Indirect Effects on Arthropod Community**  
2 **Composition and Feeding Guilds**

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12  
13 **ABSTRACT**

14 Vegetation species composition and structure are known to affect taxonomic  
15 composition and life history characteristics of arthropod communities. Soil conditions  
16 alter vegetation composition and structure, and thus soils have indirect effects on  
17 arthropods. Whilst grassland management affects soil properties, and hence  
18 vegetation, the direct effects of soil on arthropod communities within the sward is less  
19 clear. We used a long-term hay meadow experiment to assess both direct and indirect  
20 effects of various fertiliser regimes on arthropod community composition and feeding  
21 guilds. Arthropods were sampled via pitfall traps and sweep nets, then analysed using  
22 principal components and redundancy analyses (RDA) to determine relationships  
23 between soil properties, vegetation community, forage quality and arthropod  
24 community. Vegetation community composition, measured by the first vegetation  
25 principal component, was used as a constraining variable in partial RDA, to estimate  
26 direct effects of soil on the arthropods. Variance partitioning quantified the relative  
27 roles of vegetation and soil on the arthropod community. Our results indicate that  
28 available soil nitrogen and carbon-nitrogen ratios are important determinants of  
29 arthropod community composition. Once the effects of the vegetation were removed  
30 it was found the soil acidity and the available potassium altered arthropod community  
31 composition. Further research is required to determine the mechanisms by which  
32 these soil properties affect arthropod communities.

33 Keywords: Soil properties, feeding guilds, vegetation community, grasslands, nutrient  
34 additions, diversity, long-term experiments.

35

36

## 37 **INTRODUCTION**

38 Arthropods are affected by vegetative species composition, vegetation structure and  
39 grassland management practices. The plant species composition has been shown to  
40 affect the species richness and abundance of arthropods (cf. Altieri 1999; Woodcock  
41 and Pywell, 2010; Ebeling *et al.* 2012; Badenhausser *et al.* 2015; Dassou *et al.* 2016;  
42 Huelber *et al.* 2015). Schaffers *et al.* (2008) reported that plant species composition  
43 was the best predictor for an assemblage of different arthropod taxonomic groups and  
44 families. Abiotic factors affect the species composition of the vegetation, and  
45 therefore the vegetation structure, which in turn affects arthropod assemblages. For  
46 example spiders and ground beetles have been shown to be affected by ground cover  
47 (Rushton *et al.* 1989; Eyre *et al.* 1990) and butterflies prefer a sward with a non-  
48 uniform height (Jerrentrup *et al.* 2014). A heterogeneous sward will contribute  
49 generally to arthropod species richness and abundance (WallisDeVries *et al.* 2002).  
50 Another aspect of the vegetative structure is the microclimate, which has also been  
51 determined as being important in relation to arthropods. Rae *et al.* (2006) investigated  
52 its relationship with epigeal herbivores and predators and found that shorter swards  
53 generally had higher temperatures than tall swards, with a direct effect of temperature  
54 on the herbivorous community. These are all direct effects of the vegetation on  
55 arthropods.

56

57 The management regime on a grassland governs the vegetation composition and  
58 structure (Socher *et al.* 2012). A frequent hay cut, intensive grazing and high levels of  
59 fertiliser application (particularly nitrogen and phosphorus) can all result in reduced  
60 species richness (Manning *et al.* 2015), whereas a hay cut once or twice a year, light  
61 grazing and fertiliser use can result in increased species richness of vegetation and  
62 hence arthropods (Birkhofer *et al.* 2008; Mazalova *et al.* 2015). Fertiliser use alters soil  
63 nutrient status and acidity, which in turn will have a direct effect on the vegetation  
64 species composition, forage quality and structure (Crawley *et al.*, 2005), and therefore

65 an indirect effect on arthropods. However, soil may also have direct effects on  
66 arthropods once the effects of vegetation species composition have been removed.  
67 These direct effects may include soil chemical properties, such as acidity, available  
68 nitrogen and soil moisture which have been shown to have a direct effect on  
69 Collembola (Salmon *et al.* 2002; Silvertown *et al.* 2006) and soil microbial communities  
70 (Sun *et al.* 2015), but little is known concerning any direct effects of the soil on the  
71 community composition of larger arthropods commonly used in ecological studies (e.g.  
72 Coleoptera, Diptera, Hymenoptera and Hemiptera).

73

74 The aims of this paper are to: 1. Determine the effects of fertiliser regimes on the  
75 arthropod and vegetation community composition in a long-term hay meadow  
76 experiment. Specifically, to assess the differences in the community composition of the  
77 arthropods and vegetation as a result of the long-term management regimes, and to  
78 determine whether there are any differences in the main feeding guilds. 2. Assess the  
79 indirect effects of the soil characteristics on the arthropod community composition, as  
80 measured through soil properties and resultant changes in forage chemistry. 3.  
81 Determine if there are any direct effects of the fertiliser regimes and soil properties on  
82 arthropod community composition, after accounting for differences in the vegetation  
83 species composition. 4. Quantify any direct effects of the vegetation species  
84 composition and soil properties on the arthropod community.

85

86

## 87 **METHODS**

### 88 **Site description**

89 The site of Palace Leas is part of the 220 hectare mixed agriculture Cockle Park farm in  
90 Northumberland, United Kingdom, grid ref NZ202912 (55° 13' N, 1° 41' W). Palace Leas  
91 itself is a set of 14 unreplicated plots which are directly adjacent to each other, each  
92 plot is approximately 0.18ha in area (120m x 15m); see Hopkins *et al.* (2009) and Kidd  
93 *et al.* (2017) for diagrams of the experimental setup, and detailed history of the site.  
94 The experimental plots are bounded by a narrow buffer strip, hedgerow and road at  
95 the southern end, the eastern edge has a strip that contains mainly grass and thistle  
96 before a crop of barley. To the north there is another field of barley directly adjacent

97 to plot 14 and to the west a grass track then a ditch. The field adjacent to the ditch is  
98 used for light sheep grazing. The plots have had the same treatments applied every  
99 year since 1896, with the exception of plot 14 which was introduced in 1976 to reflect  
100 the amount of nutrients used at that time. The plots are cut once a year in late  
101 June/early July and the yield measured. An annual vegetation survey is also conducted  
102 in late May/early June each year. Light sheep grazing occurs in September and October  
103 each year. There is a weather station on site.

104

105 Eight plots were used (Appendix A, Table 1) that all had the same rotation system with  
106 annual organic (FYM) and/or mineral (NPK in different combinations) fertiliser  
107 application (excluded plots had 2 to 4 yearly rotations). The farmyard manure (FYM) is  
108 applied in February/early March (complete by 5<sup>th</sup> March in 2015) and is supplied by  
109 the farm for plot 1 (FYM & NPK) and plot 2 (FYM). Nitrogen is in the form of sulphate  
110 of ammonia for plot 7 (N only) and plot 13 (NPK low), whilst plot 14 (NPK high) receives  
111 ammonium nitrate and plot 1 (FYM & NPK) receives sodium nitrate. Phosphorus is in  
112 the form of triple super-phosphate, and applied to plot 1 (FYM & NPK), plot 8 (P only),  
113 plot 13 (NPK low) and plot 14 (NPK high). Potassium is applied as muriate of potash to  
114 plot 1 (FYM & NPK), plot 9 (K only), plot 13 (NPK low) and plot 14 (NPK high). The  
115 mineral fertilisers are applied late March/April; in the survey reported here it was 8<sup>th</sup>  
116 April 2015.

117

### 118 **Field sampling**

119 Arthropods were collected by both sweep nets and pitfall traps, and for the purposes  
120 of sampling, each plot was sub-divided into four quarter plots (30m x 15m). Sweep net  
121 samples were collected between the 8<sup>th</sup> and 23<sup>rd</sup> June 2015, before the hay was cut in  
122 the first week of July. During this period the mean daytime temperature was 12.98°C,  
123 mean overnight temperature 10.36°C, mean daily precipitation 0.9mm and mean  
124 sunshine 6.53 hours per day. The sweep nets used were triangular head, with an arm  
125 length of 30cm. A rectangular pattern (approximately 2m x 12m) was swept in each  
126 quarter plot, to produce a single sample in each quarter plot, providing four samples  
127 per plot. The pitfall samples were collected between the 17<sup>th</sup> and 27<sup>th</sup> August 2015,  
128 seven weeks after the hay cut. Pitfall traps are most effective in sampling epigeal

129 arthropods, whilst sweep nets would be less useful to survey the post-hay cut fauna,  
130 therefore different methods were used for the two time periods. This sampling period  
131 had a mean daytime temperature of 16.5°C, mean overnight temperature 14.3°C,  
132 mean daily precipitation 3.6mm and mean sunshine 6.13 hours per day. The traps  
133 were polypropylene cups with an opening diameter of 7.50cm and a depth of 13cm,  
134 they were filled with a saturated salt (NaCl) solution plus a small amount of detergent  
135 to break the surface tension. The pitfall traps were placed in the centre of each quarter  
136 plot, with their tops flush to the surface and left for seven days before collection, again  
137 resulting in four samples per plot.

138

139 The sward height (SH) for each plot was measured, during the sweep net sampling  
140 period, using a sward stick (Stewart *et al.* 2001); the maximum height of the sward was  
141 measured at 6 random locations within and around area sampled with the sweep nets  
142 to give a mean maximum height for each quarter plot. Vegetation species composition  
143 was sampled by percentage cover, with four 1m quadrats equally spaced, along a  
144 central transect in each plot. This data was collected two weeks prior to the arthropod  
145 sweep net sampling period. The hay is cut annually, with each quarter plot cut  
146 separately. A subsample was taken from each of the quarter plots to determine the  
147 different aspects that relate to forage quality: moisture content (MC), dry weight  
148 (DW), percentage nitrogen (N%), percentage carbon (C%) and the carbon- nitrogen  
149 ratio (CN.1) of the hay (Appendix A – Table 2). The cut was done the week following  
150 the arthropod sample collection. Soil samples were collected in September 2014 to  
151 determine, soil pH (pH), soil organic carbon content (SOC), soil carbon-nitrogen ratio  
152 (CN), soil nitrogen (SoilN), phosphorus (SoilP) and potassium (SoilK) (Appendix A –  
153 Table 3).

154

### 155 **Identification and data analysis**

156 All Insecta were identified to taxonomic family and allocated to a feeding guild, using  
157 Insects of Britain and Western Europe (Chinery, 2012). Where families contained more  
158 than one feeding guild they were allocated to the predominant (by numbers of  
159 individuals) feeding guild. The main taxon falling into this category was the Hemiptera-  
160 Miridae which contains both herbivorous and predatory species, but most Miridae

161 collected in this study belonged to herbivorous species (only a small number of  
162 specimens were found, and the second author identified all these to species). Some  
163 taxa belong to multiple guilds during their life-history and in these cases taxa were  
164 allocated to the main life-history stage collected in the study; as a result of the  
165 sampling methods used and timing of sampling, these were primarily adults. Unlike  
166 the Insecta, all Araneae belong to the same feeding guild (predators), and therefore  
167 were counted but not identified to family. Note that this decision has minor  
168 implications for the interpretation of the analyses (see Discussion). Only arthropods of  
169 approximately >2mm length were identified. A full breakdown of the arthropod taxa  
170 and allocated feeding guilds is presented in Appendix B Table 1.

171

172 Differences in the numbers of arthropod individuals as a result management were  
173 determined through linear models (one-way ANOVA) with plot treatment as the  
174 explanatory variable. Arthropod species community composition and vegetation  
175 composition (percentage cover) data were analysed using ordination techniques via  
176 principal components analysis (PCA) and redundancy analysis (RDA) and partial  
177 redundancy analysis (pRDA) to address the four main aims of the paper, using the R  
178 vegan package (Oksanen, 2015):

179 1. PCA is an unconstrained linear ordination technique, and was used to address  
180 the first aim of the study, i.e. to understand the effects of the long-term  
181 management regimes on the arthropod and vegetation community  
182 composition, and provide a convenient summary of the relative similarity in  
183 species composition. Vegetation PC1 was used as a summary variable in  
184 subsequent analyses (3 and 4 below) to partial-out the effects of vegetation  
185 species composition.

186 2. RDA is a constrained ordination methods, such that the positions of the  
187 response variable (the samples by taxa matrix) is constrained by environmental  
188 factors, so that the sample scores are linear combinations of the environmental  
189 variables. RDA was used to address the second study aim, to measure the  
190 effects of soil chemistry, and its knock-on impacts on forage properties, on the  
191 arthropods. The arthropods by samples matrix was used as the response

192 variable in the RDA, with a matrix of soil chemistry and forage quality by  
193 samples providing explanatory (constraining) variables.

194 3. pRDA differs from RDA in that partial (conditioning) explanatory variables can  
195 be incorporated into the analysis, so that the effects of other explanatory  
196 variables are easier to understand. This is particularly useful where it is  
197 suspected that some explanatory variables might have a major effect on the  
198 response, and would therefore obscure the effects of the other explanatory variables.  
199 It is well-established that vegetation species composition has a major effect on  
200 grassland arthropod communities, but a key aim of this study was to measure  
201 the effects of the fertiliser regimes after accounting for their effects on  
202 vegetation species composition. We therefore used pRDA with soil properties  
203 as standard explanatory variables, but vegetation species composition (as  
204 measured by vegetation PC1 sample scores) as a partial explanatory variable.

205 4. Variance partitioning (Borcard *et al*, 1992) uses a series of linked pRDA analyses  
206 to quantify the individual and joint effects of several explanatory variables; in  
207 this study we used it to determine the relative importance of the vegetation  
208 species composition and soil properties to quantify their relative effects on the  
209 arthropod communities.

210

211 As the raw data for the sweep net, pitfall trap and vegetation communities were  
212 dominated by a small number of families the data was standardised using the Hellinger  
213 transformation (Legendre and Gallagher, 2001) prior to principal components analysis  
214 and constrained analysis being performed. Log and square-root transformations are  
215 common alternative transformations, particularly for percentage data; however both  
216 our arthropod and vegetation community datasets were skewed by a small number of  
217 very abundant taxa, and the Hellinger transformation proved more robust in avoiding  
218 undistorted ordination diagrams. Statistical significance of explanatory variables in  
219 RDA and pRDA was tested through Monte Carlo permutation tests, using 999  
220 permutations (Oksanen, 2015).

221

222



## 223 RESULTS

### 224 Overall summary of arthropod and vegetation samples

225 A total of 3542 arthropod specimens were captured across 69 different families, with  
226 Diptera having the greatest number of individuals at 2436 (68.8%), and the largest  
227 number of families at 42. The most abundant were Scathophagidae with 1550  
228 individuals found mainly in the sweep net samples. The majority of these were the  
229 yellow dung fly, *Scathophaga stercoraria* (L. 1758). The next most numerous were the  
230 Aranae with 408 specimens and the Carabidae with 181, though both these were  
231 found mainly in the pitfall samples. A breakdown of the samples by taxonomic group  
232 and fertiliser treatment is provided in Appendix A with a complete list of all families  
233 and their functional traits in Appendix B.

234

235 The most species-rich vegetation community was in the P-only treatment (25 species)  
236 whilst the lowest was the N-only treatment (6 species). The plot with nitrogen-only  
237 additions was dominated by grass species (5) with only one forb, whereas, in the other  
238 plots the number of grass species and number of forb species were similar. The British  
239 National Vegetation Classification (NVC: Rodwell, 1991) shows both plots with FYM  
240 additions being classified as *Lolium perenne* leys and related grasslands, *Lolium*  
241 *perenne-Alopecurus pratensis* sub-community, the control and phosphorus only  
242 additions were *Cynosurus cristatus-Centaurea nigra* grassland, *Lathyrus pratensis* sub-  
243 community, both these classes are mesotrophic grasslands. The remaining plots were  
244 classified as *Festuca ovina-Agrostis capillaris-Galium saxatile* grassland, *Holcus lanatus-*  
245 *Trifolium repens* sub-community, which are calcifugous grasslands and montane  
246 communities.

247

### 248 Overall effects of fertiliser treatments on arthropod feeding guilds

249 Sweep net samples were dominated by decomposers, with 74% of 2565 individual  
250 arthropods across all plots. Of the 1906 decomposer individuals, 1070 (56%) were  
251 recorded from the two plots that received farmyard manure (FYM & NPK; FYM plots).  
252 Herbivores predominated in those plots that only received NPK (NPK low and NPK  
253 high): 185 (51%) of the total of 363 herbivores individuals. Pollinators were most  
254 abundant on the two plots that received FYM (48% of all pollinators), with the

255 remainder fairly evenly spread between the other plots. Only 36 saprophytic (fungus-  
256 feeder) individuals were collected, two thirds of these were in the two plots treated  
257 with the FYM, and none were found in those with the low NPK and high NPK plots.  
258 Linear models indicated no significant difference between the total number of  
259 individual arthropods caught by sweep netting between the plots, based on the 4  
260 samples per plot ( $F = 0.532$ ,  $p = 0.805$ ,  $df = 7$ ).

261

262 Pitfall samples were all dominated by predators, with 73% of 977 individual arthropods  
263 across all plots, but there was relatively little difference in the numbers of predators in  
264 each plot. Most herbivores were found on plots receiving only NPK treatment (low NPK  
265 and high NPK plots): 29% of all herbivore individuals. Decomposers were mainly  
266 associated with the two treatments receiving farmyard manure (FYM & NPK; FYM  
267 plots): 47% of decomposer individuals. Most parasites were found on the plot that  
268 only received phosphorus (P only): 42% of parasite individuals. Linear models  
269 indicated no significant difference between the total number of individuals caught by  
270 pitfall traps between the plots, based on the 4 samples per plot ( $F = 0.119$ ,  $p = 0.996$   $df$   
271  $= 7$ ).

272

## 273 **1. Principal components analysis to determine treatment effects on arthropod and** 274 **vegetation communities**

### 275 ***Arthropods - Sweep samples PCA results***

276 The first two axes of the principal components analysis (PCA) of the standardised  
277 sweep-net data explained 52% of the variance, with axis 1 at 34%. There is a gradient  
278 from decomposers to herbivores along the PC1 axis (Fig. 1), with the plots receiving  
279 the FYM treatments, the saprophytes and the decomposers being associated with each  
280 other. To a lesser extent, the plot receiving nitrogen-only was also associated with the  
281 saprophytes and the decomposers. The plots receiving the NPK treatments showed an  
282 association with some herbivores, especially Chrysomelidae (leaf beetles) and  
283 Delphacidae (leaf bugs). The first principal component was mainly associated with soil  
284 C:N ratio ( $R^2 = 0.497$ ) and soil N ( $R^2 = 0.496$ ), whilst the second principal component  
285 was weakly associated with soil P ( $R^2 = 0.311$ ).

286

287 FIGURE 1

288

289 ***Arthropods - Pitfall samples PCA results***

290 The first two PCA axes for the standardised pitfall trap samples explain 57% of the  
291 variance, with 38% on axis 1 (Fig. 2). Decomposers were associated primarily with low  
292 PC1 scores, and herbivores and saprophytes had higher scores, but trends along PC1  
293 were less obvious than for the sweep samples. The arthropod composition in the plots  
294 with nitrogen-only and phosphorous-only are similar, indicated by their proximity on  
295 the PCA plot, as are those in the potassium-only and NPK low. Whilst there is an  
296 association between the FYM & NPK plot and the phosphorous-only plot along the first  
297 principal component, on the second principal component the FYM & NPK plot is very  
298 different from all the other plots. The first principal component was strongly  
299 associated with soil N ( $R^2 = 0.576$ ) whilst there was a weak association between PC2  
300 and sward height ( $R^2 = 0.257$ ).

301

302 FIGURE 2

303

304 ***Vegetation – PCA results***

305 Fig. 3 summarises the vegetation species and experimental plots from the PCA (PC1  
306 65%, PC2 18%). Grass species were distributed across the PCA graph, but the forbs are  
307 concentrated around the centre of the graph, suggesting a greater degree of variation  
308 in the species composition of the grasses between the different experimental plots  
309 than for the forbs. The first principal component was strongly correlated with the  
310 available soil phosphorus ( $R^2 = 0.942$ ) and to a lesser extent with the C:N ratio ( $R^2 =$   
311  $0.796$ ) and the soil nitrogen ( $R^2 = 0.514$ ), showing a gradient from high soil phosphorus  
312 to high C:N ratio. This coincides with a gradient from those plots with FYM additions to  
313 the plots with K-only, N-only plots and the control. The second principal component  
314 had a correlation with soil pH ( $R^2 = 0.543$ ) which coincides with the phosphorus  
315 additions and the control, at high pH and the NPK high and N only additions at low pH.

316

317 FIGURE 3

318

319 **2. Constrained ordination via RDA to determine effects of environmental variables**  
320 **on the arthropod community**

321 ***RDA of sweep net samples***

322 When the sweep net samples were constrained by the soil data using RDA (fig.4a), the  
323 factors which most strongly affect the arthropod community are the C:N ratio and  
324 available soil nitrogen, though permutation tests showed no significance ( $F=1.41$ ,  
325  $p=0.181$ ,  $df=7$ ). The soil nitrogen, soil organic carbon content and the available  
326 phosphorus have similar effects, and are most closely associated with the FYM only  
327 additions and the Sepsidae family of Diptera (see Fig. 1). The C:N ratio is negatively  
328 associated to these three soil properties.

329

330 Constrained RDA with the vegetation structure and vegetation chemistry (fig.4b)  
331 shows a gradient from height to C:N ratio for the first principal component which also  
332 coincides with a FYM to NPK variance between the plots ( $F=1.174$ ,  $p=0.391$ ,  $df=7$ ). The  
333 second principal component is related to the moisture content of the vegetation with  
334 the arthropod community in the plots with the potassium and phosphorus additions  
335 preferring the drier vegetation. The factors showing the strongest effect on the  
336 arthropods are the height of the vegetation and its moisture content.

337

338 Constraining by soil properties and the PC1 scores from the plant community PCA (fig.  
339 4c), the soil C:N ratio and the vegetation PC1 scores had similar effects on the  
340 arthropod community, though the available soil nitrogen, which was negatively  
341 associated with the vegetation had an equally strong effect. The overall RDA  
342 permutation test was non-significant ( $F=1.0365$ ,  $p=0.518$   $df = 7$ ). The arthropod  
343 community in the plots with K-only and P-only treatments were associated with high  
344 vegetation PC1 scores, which itself was associated with larger amounts of vegetation  
345 species such as *Festuca rubra*, *Anthoxanthum odoratum* and *Agrostis capillaris* (refer  
346 back to Fig. 3a). Conversely, Fig 4c indicates that FYM-treated plots were associated  
347 with low vegetation PC1 scores (and higher soil N, SOC and soil P), and from Fig 3a this  
348 would suggest plants such as *Poa trivialis*, *Agrostis pratensis* and *Bromus hordeaceus*

349 were characteristic of the vegetation that was generating the largest effects on the  
350 arthropods.

351

### 352 ***RDA of pitfall samples***

353 Constrained RDA with soil data (fig. 5a) indicated the strongest soil variables for the  
354 pitfall trap samples were available soil nitrogen and the C:N ratio, i.e. the same as for  
355 sweep net samples, and also non-significant ( $F=1.162$ ,  $p=0.428$ ,  $df=7$ ). A high C:N ratio  
356 is closely associated with the arthropod community from the plot with NPK low  
357 additions, and to a slightly lesser extent the plot with the K only additions.

358

359 When constrained by the vegetation structure and vegetation chemistry, the C:N ratio  
360 and dry weight of the vegetation are the strongest factors (fig. 5b). The community in  
361 the NPK low and K only plots are aligned with the higher C:N ratio and prefer the lower  
362 nitrogen content ( $F=0.823$ ,  $p=0.760$ ,  $df=7$ ). The dry weight of the vegetation is the  
363 strongest explanatory variable, with a gradient from the control plot to FYM & NPK.

364

365 Constraining by the vegetation species composition (vegetation PC1 score) and soil  
366 properties on the epigeal community (fig. 5c) showed similar patterns to that from the  
367 sweep samples, with both vegetation PC1 and soil C:N ratio being highly correlated  
368 and important predictors ( $F=1.9422$ ,  $p=0.077$ ,  $df = 7$ ). High C:N ratio and vegetation  
369 PC1 scores were associated with sites receiving NPK (high or low) or K-only; again,  
370 cross-reference back to Fig. 3a suggests the epigeal arthropods in these samples are  
371 from vegetation dominated by *F. rubra*, *A. odoratum* and *A. capillaris*.

372

### 373 **3. Partial constrained ordination via pRDA to determine effects of environmental 374 variables on the arthropod community after accounting for vegetation composition**

#### 375 ***pRDA of sweep net samples***

376 The effects of the soil properties on the sweep net arthropods, once the effects of the  
377 vegetation community have been removed by using vegetation PC1 as a partial  
378 variable (Fig. 4d) indicated that the soil K and soil pH had the largest effects on the  
379 arthropod community composition. The effects of soil N and C:N ratio on the

380 arthropods was reduced once the vegetation community composition had been  
381 partialled-out (compare Figs. 4a and 4d).

382

383 The constraining variable soil K was, as might be expected, associated with the  
384 experimental plot receiving the K-only treatment (Fig. 4d). The arthropods in this  
385 treatment were predominantly characterised by an abundance of Curculionidae (see  
386 Fig. 1). High soil pH was related particularly to the plots receiving FYM & NPK, and P-  
387 only, and conversely low pH with the NPK-low plot (by extending the pH arrow in Fig  
388 4d backwards). Cross-reference with Fig. 1 suggests that arthropods characteristic of  
389 alkaline conditions include Chloropidae, Scathophagidae and Sciaridae. High soil N  
390 and SOC were characteristic of the N-only treatment; this is not characterised by any  
391 particular group of arthropods.

392

393 FIGURE 4

394

#### 395 ***pRDA of pitfall net samples***

396 Community composition of arthropods collected by pitfall trap is most strongly  
397 affected by the soil K, N and pH (fig. 5d) once the effects of the vegetation had been  
398 removed via partial RDA (vegetation PC1 scores). The control plot was associated with  
399 high soil pH and N, which from Fig. 2 suggests Lonchopteridae, and Cynipidae are  
400 affected by available nitrogen. As might be expected, soil K was strongly positively  
401 associated with the K-only treatment. The FYM & NPK treatment point is  
402 approximately equidistant from potassium and soil pH arrows in Fig 5d, suggesting that  
403 both constraining variables are important. Cross-reference with Fig. 2 suggests  
404 Chrysomelidae were particularly associated with this treatment plot.

405

406 FIGURE 5

407

#### 408 **4. Variance partitioning via pRDA to quantify direct and indirect effects**

409 The main direct effect identified by variance partitioning on the species composition of  
410 the sweep-net arthropods was the vegetation species composition, at 27.1%. In  
411 contrast, the vegetation had a smaller effect on pitfall (epigeal) arthropods caught via

412 pitfall traps (13.8%), whilst the soil chemistry had a larger effect on epigeal arthropods  
413 in comparison to the sweep net arthropods (11.6% vs 8.2%, Table 1). The joint  
414 amount of variation explained by both the vegetation and the soil was relatively low at  
415 approximately 4% in both cases. However, irrespective of sampling method, over 50%  
416 of the variation in the arthropod community was unexplained by either soil chemistry  
417 or vegetation species composition.

418

419

420 TABLE 1

421

## 422 **DISCUSSION**

423 This study has investigated the relationships between the vegetation species  
424 composition, structure, forage quality, soil properties, and the arthropod community,  
425 in a long-term hay meadow experiment. Our results demonstrate that the  
426 management has both direct (changes in soil properties) and indirect effects (mediated  
427 through the vegetation) on the arthropod community composition.

428

429 The vegetation species community had a strong relationship with the soil in terms of  
430 available phosphorus, soil acidity and the carbon-nitrogen ratio, and the arthropod  
431 community had a strong relationship with nitrogen, both in terms of the available  
432 nitrogen in the soil and the nitrogen content of the vegetation. This agrees with  
433 previous studies by La Pierre and Smith (2016) who suggested that nutrient additions,  
434 especially nitrogen, increased the abundance of herbivorous insects in the mesic  
435 prairies of North America. The arthropod herbivores, especially Chrysomelidae and  
436 Delphacidae, are more abundant where the C:N ratios are higher and therefore the  
437 relative amount of nitrogen is lower. Whilst this result does not agree with the results  
438 of many previous studies (cf. Wang *et al.* 2006,) there have been some instances  
439 where nitrogen additions and therefore a lower C:N ratio have been shown to either  
440 reduce or make no change to herbivore abundance (Bethke *et al.* 1998, Casey and  
441 Raupp, 1999). Rae *et al.* (2006) found that the microclimate in the sward, especially  
442 soil and air temperature, can have a direct effect on the herbivore community; in our

443 study treatments with high N resulted in greater vegetation growth, which may have  
444 affected the microclimate. The relationship between nitrogen and the arthropod  
445 community is due to the effects that the nitrogen availability and C:N ratios have on  
446 the vegetation and is therefore an indirect relationship.

447

448 The long term addition of nitrogen-only fertilisers produced a similar arthropod  
449 community structure to that of the plots which had farmyard manure additions, even  
450 though the vegetation showed more similarity with the plots that had the NPK  
451 additions. This may be because a layer of organic matter, 6cm thick, had developed  
452 over the top of the soil at Palace Leas as a result of the low acidity caused by the  
453 nitrogen additions (Shiel and Rimmer, 1984). So the results for the nitrogen-only plot  
454 show that it is the resultant soil properties that have influenced this arthropod  
455 community more than the vegetation. Short term nitrogen only fertiliser use may not  
456 have the same results.

457

458 The biggest differences in the arthropod communities were between the plots with  
459 FYM treatment and those with NPK treatments. Plots receiving FYM & NPK nutrient  
460 additions showed many similarities with those that only received FYM and was very  
461 different to the mineral fertiliser (NPK) plots, suggesting that farmyard manure may be  
462 a dominant factor determining the species composition of arthropod communities,  
463 especially decomposers and saprophytes. Whilst the vegetation species composition  
464 was also affected by FYM application, the differences between the plots were not as  
465 obvious as for the arthropod communities.

466

467 Most previous studies have investigated the joint effects of vegetation and  
468 environmental factors on arthropod communities (Birkhofer *et al.* 2008; Crist and  
469 Peters, 2014; Manning *et al.* 2015). In contrast, we have attempted to identify the  
470 environmental factors that have both direct and indirect effects on the arthropods,  
471 depending on how they are mediated by the vegetation. Removal of the vegetation  
472 community composition, via partial RDA, indicated that arthropod community  
473 composition, collected by sweep net, was directly affected by soil pH and the available  
474 potassium in the soil. The epigeal community was also affected by the available



475 nitrogen in the soil. These environmental factors had not shown any importance until  
476 the vegetation species community data had been removed, suggesting that the effects  
477 of the carbon-nitrogen ratio, and soil organic carbon on the arthropod community  
478 were indirect and due to the effects that these factors had on the vegetation, whereas  
479 soil pH, available potassium and for the epigael community, available nitrogen, do  
480 have direct effects. This may be in relation to the fact that they somehow affect the  
481 microclimate within the sward, especially with the volatilisation of nitrogen (Brown *et*  
482 *al.* 2010). The main direct and indirect factors are summarised in Fig. 6.

483

484 FIGURE 6

485

486 When variance partitioning was applied the vegetation species composition had a  
487 larger influence on the arthropod community than the soil properties, and the soil  
488 properties have more of an effect on the epigael community than the sweep net  
489 community. These results are in agreement with many other studies (Curry 1994;  
490 Harte *et al.* 1996; Schaffers *et al.* 2008). However, there is a high proportion which  
491 remains unexplained. This may be due to the structure of the vegetation and the  
492 microclimate created by it. These results do however, show a direct effect from the  
493 resultant soil properties due to the different fertiliser regimes, with soil nitrogen being  
494 the most influential factor.

495

496 A novel aspect of this research is that the direct effects of the soil conditions on  
497 arthropod community composition have been identified, as well as the readily  
498 measured vegetation effects. Whilst the latter has been reported widely in the  
499 scientific literature (*cf.* Schaffers *et al.* 2008; Woodcock and Pywell, 2010), there are  
500 relatively few studies of the former. The exact causal mechanisms for the soil directly  
501 affecting the arthropod community composition cannot be ascertained directly in the  
502 current study, but they might be associated with factors such as microclimate (e.g.  
503 differences soil moisture) and nutritional value of plants (e.g. available nitrogen to  
504 arthropod herbivores).

505

506 One weakness of this study relates to the fundamental design of the experiment when  
507 it was originally create in the late 19<sup>th</sup> Century. None of the treatments are replicated  
508 in separate plots, and so technically the four sets of samples we have collected within  
509 each plot could be viewed as 'pseudo-replicates'. The design of the experiment would  
510 have been better if there had been replicate plots and a block design, but such  
511 concepts in statistical theory and experimental design were not fully understood at the  
512 time. Nevertheless, an advantage of utilising an experiment such as this is that it  
513 provides a unique opportunity to gain insights into the effects of treatments when  
514 applied continuously over long time-periods. Vegetation changes that arise from  
515 short-term experiments of only a few years duration differ significantly from long-term  
516 studies. For example, Kidd *et al.* (2017) in a parallel study of the soil and vegetation at  
517 Palace Leas demonstrated that plots receiving only artificial fertilisers, with the  
518 exception of the high-NPK treatment, became dominated by lower yielding, acid-  
519 tolerant species. Our results indicate that such long-term changes in the soil and  
520 vegetation will inevitably have knock-on effects on the arthropods.

521

522 A second limitation is that there is no single sampling method that is suitable for foliar  
523 and epigeal arthropods, and both techniques have biases, are affected by  
524 meteorological conditions, vegetation density etc. It must be recognised that  
525 arthropods sampled from within one plot may have originated from neighbouring  
526 plots, and this will make it more difficult to detect true treatment effects on the  
527 taxonomic composition of the arthropods. Finally, the interaction between vegetation  
528 biochemistry, structure and arthropods is complex, and affected by feeding type. For  
529 example, it has long been established that higher plant nitrogen generally increases  
530 vulnerability to arthropod herbivores (McNeil and Southwood, 1978), but the role of  
531 predatory/parasitic species is more complex, as they may detect the arthropod prey  
532 directly, or volatiles (including induced-defence chemicals) released by the vegetation  
533 when under attack by herbivores (Poelman and Dicke, 2014). Many arthropod guilds  
534 are affected by changes in vegetation structure (Woodcock and Pywell, 2010) The  
535 different management regimes affected vegetation structure, which has major impacts  
536 on some groups of arthropods, especially for some of the larger species within the  
537 Araneae (Mcnett and Rypstra, 2000) .

538

539 In summary, the soil properties and the vegetation have indirect effects, arising from  
540 the different fertiliser regimes, on the species composition of the arthropod  
541 community, whilst the addition of farmyard manure provides direct effects on the  
542 decomposers and saprophytes. This study has been done in a single year, on a long-  
543 term experiment, and it should be emphasised that it may take many years before  
544 these separate mechanisms can be identified.

545

546

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552

#### 553 **REFERENCES**

554 Altieri MA, 1999. The ecological role of biodiversity in agroecosystems. *Agr Ecosyst*  
555 *Environ*, 74, 19-31.

556 Badenhauer I, Gross N, Cordeau S, Bruneteau, L, Vandier M, 2015. Enhancing  
557 grasshopper (Orthoptera: Acrididae) communities in sown margin strips: the role of  
558 plant diversity and identity. *Arthropod-Plant Inte*, 9, 333-346.

559 Bethke J, Redak R, Schuch UK, 1998. Melon aphid performance on chrysanthemum as  
560 mediated by cultivar, and differential levels of fertilization and irrigation. *Entomol Exp*  
561 *Appl*. 88, 41-47.

562 Birkhofer K, Fliessbach A, Wise D, Scheu S, 2008. Generalist predators in organically  
563 and conventionally managed grass-clover fields: implications for conservation  
564 biological control. *Ann Appl Biol*, 153, 271-280.

565 Brown J, Scholtz C, Janeau J, Grellier S, Podwojewski P, 2010. Dung beetles  
566 (Coleoptera: Scarabaeidae) can improve soil hydrological properties. *Appl Soil Ecol*, 46,  
567 9-16.

568 Borcard D, Legendre P, Drapeau P, 1992. Partialling Out the Spatial Component of  
569 Ecological Variation. *Ecology*, 73, 1045-1055.

570 Casey C, Raupp M, 1999. Supplemental nitrogen fertilization of containerized azalea  
571 does not affect performance of azalea lace bug (Heteroptera: Tingidae). *Environ*  
572 *Entomol*, 28, 998-1003.

573 Chinery M, 2012. *Insects of Britain and Western Europe*. Bloomsbury Publishing,  
574 London.

575 Crawley MJ, Johnston AE, Silvertown J, Dodd M, Mazancourt C de, Heard MS, Henman  
576 DF, Edwards GR, 2005. Determinants of Species Richness in the Park Grass Experiment.  
577 *Am Nat*, 165, 179-192.

578 Crist TO, Peters VE, 2014. Landscape and Local Controls of Insect Biodiversity in  
579 Conservation Grasslands: Implications for the Conservation of Ecosystem Service  
580 Providers in Agricultural Environments. *Land*, 3, 693-718.

581 Curry JP, 1994. *Grassland invertebrates: ecology, influence on soil fertility, and effects*  
582 *on plant growth*. Chapman & Hall, London.

583 Dassou AG, Depigny S, Canard E, Vinatier F, Carval D, Tixier P, 2016. Contrasting effects  
584 of plant diversity across arthropod trophic groups in plantain-based agroecosystems.  
585 *Basic Appl Ecol*, 17, 11-20.

586 Ebeling A, Klein A-M, Weisser W, Tschardt T, 2012. Multitrophic effects of  
587 experimental changes in plant diversity on cavity- nesting bees, wasps, and their  
588 parasitoids. *Oecologia*, 169, 453-65.

589 Eyre MD, Luff ML, Rushton SP, 1990. The Ground Beetle (Coleoptera, Carabidae) Fauna  
590 of Intensively Managed Agricultural Grasslands in Northern England and Southern  
591 Scotland. *Pedobiologia*, 34, 11-18.

592 Harte J, Rawa A, Price V, 1996. Effects of manipulated soil microclimate on mesofaunal  
593 biomass and diversity. *Soil Biol Biochem*, 28, 313-322.

594 Hopkins DW, Waite IS, McNicol JW, Poulton PR, Macdonald AJ, O'Donnell AG, 2009.  
595 Soil organic carbon contents in long-term experimental grassland plots in the UK  
596 (Palace Leas and Park Grass) have not changed consistently in recent decades. *Global*  
597 *Change Biol*. 15, 1739-1754.

598 Huelber K, Haider JA, Hager TE, Dullinger S, Fiedler K, 2015. Insect herbivory in alpine  
599 grasslands is constrained by community and host traits. *J Veg Sci*, 26, 663-673.

600 Jerrentrup JS, Wrage-Mönnig N, Röver K-U, Isselstein J, 2014. Grazing intensity affects  
 601 insect diversity via sward structure and heterogeneity in a long-term experiment. *J*  
 602 *Appl Ecol*, 51, 968-977. Kidd J, Manning P, Simkin J, Peacock S, Stockdale E, 2017.  
 603 Impacts of 120 years of fertilizer addition on a temperate grassland ecosystem. *PLOS*  
 604 *One*, 12(3): e0174632. <https://doi.org/10.1371/journal.pone.0174632>  
 605 La Pierre KJ, Smith MD, 2016. Soil nutrient additions increase invertebrate herbivore  
 606 abundances, but not herbivory, across three grassland systems. *Oecologia*, 180, 485-  
 607 497.  
 608 Legendre P, Gallagher ED, 2001. Ecologically meaningful transformations for ordination  
 609 species data. *Oecologia*, 129, 271-280.  
 610 Manning P, Gossner MM, Bossdorf O, Allan E, Zhang YY, Prati D, Bluthgen N, Boch S,  
 611 Bohm S, Borschig C, Holzel N, Jung K, Klaus VH, Klein AM, Kleinebecker T, Krauss J,  
 612 Lange M, Muller J, Pasalic E, Socher SA, Tschapka M, Turke M, Weiner C, Werner M,  
 613 Gockel S, Hemp A, Renner SC, Wells K, Buscot F, Kalko EKV, Linsenmair KE, Weisser  
 614 WW, Fischer M, 2015. Grassland management intensification weakens the associations  
 615 among the diversities of multiple plant and animal taxa. *Ecology*, 96, 1492-1501.  
 616 Mazalova M, Sipos J, Rada S, Kasak J, Sarapatka B, Kuras T, 2015. Responses of  
 617 grassland arthropods to various biodiversity-friendly management practices: Is there a  
 618 compromise? *Eur J Entomol*, 112, 734-746.  
 619 McNeil S, Southwood TRE 1978. The role of nitrogen in the development of insect-  
 620 plant relationships. In: Harborne JB (ed) *Aspects of plant and animal co-evolution*,  
 621 Academic Press, London, 77-98.  
 622 Oksanen J, Blanchet FG, Kindt R, Legendre P, Minchin PR, O'Hara RB, Simpson GL,  
 623 Solymos P, Henry M, Stevens H, Wagner H, 2015. *Vegan: Community Ecology Package*.  
 624 R package version 2.2-1. URL <http://CRAN.R-project.org/package=vegan>.  
 625 Mcnett BJ, Rypstra AN 2000. Habitat selection in a large orb-weaving spider:  
 626 vegetational complexity determines site selection and distribution. *Ecol Entomol*, 25,  
 627 423-432.  
 628 Poelman EH, Dicke M 2014. Plant-mediated interactions among insects within a  
 629 community ecological perspective. *Ann Plant Rev*, 47, 309-338.  
 630 R Core Team, 2013. *R: A language and environment for statistical computing*. R  
 631 Foundation for Statistical Computing, Vienna, Austria. URL <http://www.R-project.org/>.

632 Rae DA, Armbruster WS, Edwards ME, Svengård-Barre M, 2006. Influence of  
633 microclimate and species interactions on the composition of plant and invertebrate  
634 communities in alpine northern Norway. *Acta Oecol*, 29, 266-282.

635 Rodwell JS, 1991. *British plant communities*. Cambridge University Press, Cambridge.

636 Rushton SP, Luff ML, Eyre MD, 1989. Effects of Pasture Improvement and Management  
637 on the Ground Beetle and Spider Communities of Upland Grasslands. *J Appl Ecol*, 26,  
638 489-503.

639 Salmon S, Ponge J, Van Straalen N, 2002. Ionic identity of pore water influences pH  
640 preference in Collembola. *Soil Biol Biochem*, 34, 1663-1667.

641 Schaffers AP, Raemakers IP, Sykora KV, Ter Braak CJF, 2008. Arthropod assemblages  
642 are best predicted by plant species composition. *Ecology*, 89, 782-794.

643 Shiel RS, Rimmer DL, 1984. Changes in soil structure and biological activity on some  
644 meadow hay plots at Cockle Park, Northumberland. *Plant Soil*, 76, 349-356.

645 Silvertown J, Poulton P, Johnston E, Edwards G, Heard M, Biss PM, 2006. The Park  
646 Grass Experiment 1856-2006: It's Contribution to Ecology. *J Ecol*, 94, 801-814.

647 Socher SA, Prati D, Boch S, Muller J, Klaus V, Holzner N, Fischer M, 2012. Direct and  
648 productivity-mediated indirect effects of fertilization, mowing and grazing on grassland  
649 species richness. *J Ecol*, 100, 1391-1399.

650 Stewart KEJ, Bourn NAD, Thomas JA, 2001. An evaluation of three quick methods  
651 commonly used to assess sward height in ecology. *J Appl Ecol*, 38, 1148-1154.

652 Sun R, Guo X, Wang D, Chu H, 2015. Effects of long- term application of chemical and  
653 organic fertilizers on the abundance of microbial communities involved in the nitrogen  
654 cycle. *Appl Soil Ecol*, 95, 171-178.

655 WallisDeVries MF, Poschlod P, Willems JH, 2002. Challenges for the conservation of  
656 calcareous grasslands in northwestern Europe: integrating the requirements of flora  
657 and fauna. *Biol Conserv*, 104, 265-273.

658 Wang J, Tsai J, Broschat T, 2006. Effect of nitrogen fertilization of corn on the  
659 development, survivorship, fecundity and body weight of *Peregrinus maidis* (Hom.  
660 Delphacidae). *J Appl Entomol*, 130, 20-25.

661 Woodcock BA, Pywell R, 2010. Effects of vegetation structure and floristic diversity on  
662 detritivore, herbivore and predatory invertebrates within calcareous grasslands.  
663 *Biodivers Conserv*, 19, 81-95.